

1 – INTRODUCTION

 Sediment compartments have been increasingly used in quality evaluation studies in aquatic ecosystems as they record historic conditions of anthropic activity influences in these environments which are not always detectable in water compartments. The interaction between the sediment compartment and the water column can, in itself, be a source of contamination and the interaction can occur through the revolving sediment because of the increased flow due to rainfall or activity that interfere with the dynamics of the riverbed (CETESB 2011). The sediment, from the recycling of material and energy flow point of view, is one of the most important compartments of aquatic ecosystems. Biological, physical and/or chemical processes occurring therein, influencing the metabolism of the entire system, since the benthic organisms, through ichthyofauna and algae.

 Metals of anthropic origin upon entering in an aquatic system can be predominantly in their 54 dissolved form $(M(aq)^{+x})$ and depending on their characteristics, can remain dissolved or form organic and/or inorganic complexes which, in turn can remain in a dissolved state, form colloids or be absorbed by suspended particles in the water body and be deposited in the sediment (Silvério 2003, Cánovas et al., 2012).

 One of the big questions in environmental evaluations is how much an ecosystem is contaminated by a given metal. Many times the anthropic enrichment of a given element may cause great concern even though the high concentration may not result in toxic effects due to the fact this metal be in a form not readily available to the biota, either as sulfide or highly complexed with the organic matter. Another important aspect to take into consideration is if the internationally accepted sediment quality guide values used (for example, PEL, TEL or crust average) are adequate for tropical conditions (Quináglia 2012; Gomes et al. 2009), because if not, these can lead to interpretation errors of metal evaluations in aquatic bodies.

 However, when dealing with artificial reservoirs this issue becomes a complex task, especially in function of their allochthonous characteristics. The sediments of reservoirs or dams operate as true deposits of particulate materials from the rivers that form them. Artificial reservoirs have high sediment rates when compared to those of the rivers that flow into them which are lotic system and the reservoirs lentic system. This difference influences the local sedimentation rates. Furthermore, depending on the location, the alterations that occur in the sediment may be characterized in relatively short periods of time, such as little as five years or less and a 10 cm layer of recent sediment. As a consequence, artificial reservoirs are little affected by diagenetic processes and provide a preserved historic record, from a lithological point of view, of the metal concentrations in their deeper sediment profiles (Audry et. al. 2004).

 Metal evaluations in sediment profiles (Audry et al. 2004, Gomes et al. 2009) to obtain the basal levels of the sediments, mainly in tropical reservoirs where the dynamics of temperature and consequently weathering are greater (Albarede 2011) are a key point to ascertain whether or not an environment presents anthropic contamination or merely a lithological enrichment. In order to best establish this difference both the Enrichment Factor (EF) and Geoaccumulation Index (Igeo) are used in the present study (Sutherland 2000, Audry et al. 2004, Luiz-Silva et al. 2008, Franklin et al. 2012a).

 The Rio Grande Reservoir is located in the and covers the São Bernardo do Campo, Rio Grande da Serra, Santo André and Ribeirão Pires municipalities in the state of São Paulo in Brazil. This reservoir supplies water for around 1.8 million inhabitants. It is also used for leisure and recreation by the city´s population and its fish is consumed by inhabitants of the surrounding regions.

 This reservoir was built in the 1920's and initially was one of the branches of the Billings Reservoir. In 1982, the Rio Grande Reservoir was separated from the already highly polluted Billings Reservoir in an attempt to preserve its water quality. This separation eliminated the entrance of polluted waters from the city of São Paulo (Franklin et al. 2012a).

 The objective of this study was to evaluate two sediment profiles from the Rio Grande Reservoir in order to determine basal values of some trace, metals and rare earth elements as well as to 92 determine their level of anthropic pollution. The dates and sedimentation rates by ^{210}Pb method in one of the sediment profiles were also determined.

2 – MATERIALS AND METHODS

 Two sediment cores (0.40 and 0.36 m) were collected in November 2011 in front of the water catchment point of the reservoir for the public water system, coordinates S 23°46'18''and W 46°31'35'', as showed in Figure 1. The first sediment profile was sliced every 5 cm, air dried at 20- 100 25°C in a clean recipient, and separated in two aliquots: one for grain size determination and another for Hg and other metals by ICP-OES (Inductively Coupled Plasma – Optical Emission spectrometry) and GF AAS (Graphite Furnace Atomic Absorption Spectrometry). This aliquot was passed through a 2.00 mm sieve, ground in a mortar and then homogenized before analysis. The second sediment profile was sliced every 2cm, air dried at 20-25ºC in a clean recipient, passed through a 2.00 mm sieve, ground in a mortar and then homogenized before INAA and dating analysis.

107 **Figure 1 – Sampling Point Location**

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109 **2.1 – Instrumental neutron activation analysis**

 Instrumental neutron activation analysis (INAA) has been extensively employed in geochemical studies due to the possibility of quantifying many elements at the same time with excellent precision and accuracy and without the need of previous digestion processes. The detection limits varied from 0.01 to 1 mg kg⁻¹ for most elements.

 For the multi-elemental analysis, approximately 150 mg of sediment (duplicate samples) and reference materials were accurately weighed and sealed in pre-cleaned double polyethylene bags, for irradiation. Single and multi-element synthetic standards were prepared using pipettes with convenient aliquots of standard solutions onto small sheets of Whatman No. 41 filter paper. Sediment samples, 118 reference materials and synthetic standards were irradiated for 8 h under a thermal neutron flux of 10^{12} μ n cm⁻² s⁻¹ in the IEA-R1 nuclear research reactor at IPEN. Two series of counting were made: the first, after one week decay and the second, after 15-20 days. Gamma spectrometry was performed using a 121 Canberra HPGe detector and associated electronics, with a resolution of 0.88 and 1.90 keV for ${}^{57}Co$ 122 and 60 Co respectively. The elements determined by using this methodology were As, Ba, Br, Co, Cr, Cs, Fe, Hf, Na, Rb, Sb, Sc, Ta, Th, U, Zn and the rare earths Ce, Eu, La, Lu, Nd, Sm, Tb and Yb. For methodology validation regarding precision and accuracy reference materials SL-1 (Lake Sediment – IAEA), BEN-Basalt-IWG-GIT and Soil 5 (IAEA) were used.

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2.2 – ²¹⁰ 127 **Pb Dating Methodology**

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The dates and sedimentation rates were determined by ^{210}Pb method using the measurement of 130 the radionuclides 226 Ra and 210 Pb in each slice of the core. The samples of sediments after drying at 20-131 25 °C were passed through a 2.00 mm sieve, ground in a mortar, homogenized and dissolved in 132 mineral acids, $(HNO_3 \text{conc.}$ and $HF 40\%)$, and $H_2O_2 30\%$ in microwave digestor and submitted to the radiochemistry procedure for the determination of Ra and Pb. This procedure consists in an initial 134 precipitation of Ra and Pb with H_2SO_4 3M, dissolution of the precipitate with nitrilo-tri-acetic acid at 135 basic pH, precipitation of $Ba(^{226}Ra)SO_4$ with ammonium sulfate and precipitation of $^{210}PbCrO_4$ with 136 sodium chromate 30%. The 226 Ra concentration was determined by gross alpha counting of the 137 precipitate of $Ba^{(226}Ra)SO_4$ and the ²¹⁰Pb concentration through its decay product ²¹⁰Bi, by measuring 138 the gross beta activity of the precipitate of $^{210}PbCrO₄$. The analyses were performed in duplicate and both radionuclides were determined in a low background gas flow proportional detector (Damatto 2009).

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2.3 – Metals determined by ICPOES, GF AAS and Direct Mercury Analyzer

 The sediment samples were previously digested according EPA 3051 procedure (EPA 2007), 145 with HNO₃ and HCl in microwave oven. The metals Cu, Ni, and Mn were determined by ICPOES and Cd and Pb, by GF AAS. Hg was determined by Direct Mercury Analyser. These analyses were performed at Inorganic Chemistry Laboratory of the Environmental Agency of São Paulo State (CETESB). The analyses were performed in duplicate.

 Direct analysis of Hg has the advantage of dispensing steps of sample preparation and is applicable to virtually some matrix. This technique effectively determines the Total Mercury because 151 it includes organic forms, not always possible to be destroyed in a $HNO₃$ digestion and reduced with stannous chloride. This technique includes determining species Hg volatilizables under the conditions of the method.

 There are three sequential steps that comprise this analytical technique, thermal decomposition, in which the sample undergoes a drying process and combustion (approx. 650 °C) under constant flow 156 of oxygen gas (O_2) . These combustion products are loaded by O_2 to a catalytic section which will retain possible interferences in the determination of Hg. In the amalgamation, the remaining gaseous species in the process of thermal decomposition are loaded to an amalgamator. After a retention time of Hg vapor, the amalgamator is heated to 850 °C to release the trapped Hg, which volatilizes to be 160 loaded by the O_2 flow to the cell atomic absorption, where the transient Hg vapor will absorb the radiation from an Hg vapor lamp emitting a narrow band of wavelengths of 254 nm.

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2.4 – Sediment Grain Size determination

 Sediment grain size determination was performed according to standard CETESB L6.160 (CETESB 1995). This analysis is based on the principles of sieving and sedimentation guided by Wentworth scale, which is based on the average velocities of particles in an aqueous medium. According to this standard method, the sand fraction consists of particles with dimensions between 2.0 and 0.063 mm, coarse sand being comprised between the diameters 1.0 to 2.0 mm. The silts are particles with size between 0.063 and 0.004 mm, clays with a diameter smaller than 0.004 mm.

2.5 – EF and *Igeo*

 The Enrichment Factor (EF) is a factor which allows the evaluation of the enrichment of a given element through the normalization of another element considered stable and fixed in the environment and then compared to reference concentrations taking into account a normalizing element. This tool was proposed in 1979 by Buat-Menard (Loska et al. (1997, 2003 e 2004), Szefer & Skwarzec 1988) and has been applied in several regions of the world to evaluate anthropic enrichment of certain elements. The formula used for this calculation follows (equation 1):

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EF = \frac{\left(\frac{Me}{X}\right)_{loc}}{\left(\frac{Me}{X}\right)_{ref}}(1)
$$

 Where EF -enrichment factor Me - metal or element analyzed

X -metal or element normalizer

Loc -study location

Ref -reference values

 Many elements can be used as normalizing elements such as Sc, Fe, Al, Ti, Y and Li (Loska et. al. 2003, Sutherland 2000, Lin et. al. 2008, Dias & Prudêncio 2008, Hernandez et al. 2003). The present study used Sc as the normalizing element.

188 As an enrichment evaluation criteria, some authors accept values between $0.5 \leq EF \leq 1.5$ as 189 indicators that an element is not enriched, while values ≥ 1.5 indicate element enrichment (Zhang & Liu, 2002). However, for Hernandez et al. (2003), only when the EF >2.0 can be considered of anthropic origin. Sutherland (2000), after justifying the absence or lack of criteria to define a level of pollution for the EF, proposed five enrichment categories as shown in Table 1.

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Table 1 -EF Categories (Sutherland 2000)

 The Geoaccumulation Index (*Igeo*) is a geochemical tool used to evaluate the contamination of a given location. The *Igeo* can be applied to organic and inorganic substances to evaluate

198 contamination by way of comparing values of samples of pre-industrial locations or locations with no 199 impact of the substances of interest. This index has been used for at least thirty years and is calculated 200 using equation 2.

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I_{geo} = \log_{2} \left(\frac{C_n}{1.5 \cdot B_n} \right)
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 (2)

202 Where;

203 Log2 -base 2 logarithms.

204 C_n -element concentration at evaluated location

 B_n -element concentration in the reference values used(background)

206 A factor of 1.5 for B_n according to several authors (Audry et al. 2004, Cukrov et al. 2011, Gomes et al. 2009, Loska et al.(1997, 2003 and 2004), Lin et al. 2008, Mortatti & Probst 2010; Rubio et. al. 2000) is applied to correct small fluctuations of lithogenic origins or even small anthropic 209 influences in relation to background values. In the same manner of the EF, the B_n variable represents the reference values (background) that should be used to represent element basal concentration. The same considerations for EF are used for this calculation. Several authors use earth crust values and still others prefer local values believing these to be more representative.

- 213 The *Igeo* Index is associated to a qualitative pollution intensity scale shown in Table 2.
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217 Both geochemical tools (EF and *IGeo*) for metal contamination studies are widely used by 218 various researchers using different sources as background references.

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220 **3 – RESULTS AND DISCUSSION**

222 **3.1 – Granulometric composition of the sediments**

223 Table 3 shows the results for the grain size determination of the sediment slices.

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226 Particle size determination indicates that the sediment is predominantly silty-clayey throughout 227 the profile, and characterized by a constant input of fine grained material.

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3.2 – ²¹⁰ 229 **Pb profile and** s**edimentation rates**

230 Isotopic ages for the core are presented in Figure 2. The ^{210}Pb profile depicts a linear trend 231 (r=0.9687) from level 0 cm (year 2011) to level 36 cm (year 1922), with a mean sedimentation rate of 232 0.40 cm y^{-1} . This ²¹⁰Pb profile can be divided in two trends: from level 0 cm (year 2011) to level 12cm 233 (year 1964) indicating an average accumulation rate of 0.26 cm y⁻¹ and from level 12 cm to level 36 cm 234 (year 1922) indicating an average accumulation rate of 0.57 cm $y⁻¹$. Lower rates were related to the 235 period prior to the water dam, when the loading of the sediments was stabilized. Higher rates were 236 related to rainy seasons and urban expansion in the areas surrounding the reservoir. The values 237 obtained at the latest slice were considered the background of the reservoir. This slice corresponded to 238 sediment age of the 1920's.

Figure 2 - ²¹⁰ Pb profile and sedimentation rates

3.3 – Multielemental determinations by INAA

 The precision and accuracy of the INAA analytical methodology was verified by reference 244 materials analysis and Z value calculation was done according to Bode (1996). If $|Z| < 3$, the individual result of the control sample (reference material) lies on the 99% confidence interval of the target value. For the reference materials analyzed in the present study all the results were in the 247 interval range of $|Z|$ <3, indicating good precision and accuracy of the INAA technique (Figure 3).

 Table 4 presents the results and respective uncertainties (calculated by error propagation) obtained by INAA. Tables 5 and 6 present EF and Igeo results, respectively.

 EF was calculated using Sc as the normalizing element and the results obtained for the last layer of the 0.40 cm sampled profile as reference values (also used as reference values for *Igeo* calculations) that reflects the concentration of these elements during the 1920's when industrialization in the São Paulo metropolitan region was still incipient (Table 5).

 EF results indicated that most of the elements did not present enrichment as did *Igeo* results for the elements As, Ba, Ca, Cr, Co, Cs, Fe, Sm, Eu, Lu, Rb, Th, Nd and Ta which presented values always *Igeo* < 0.0 and EF < 2.0 for the entire profile evaluated (Tables 5 and 6). According to Sutherland (2000), EF < 2.0 cannot be considered as anthropic enrichment.

Table 4 – Continuation – Results obtained for the sediment samples by INAA along the profile (mg kg⁻¹)and dating of the sediment slices 271 **(continuation)**

Depth	Year	Na.	Nd	Rb	Sb	Sc	Sm	Ta	Th	U	Yb	Zn
$0 - 2$	2011 - 2004	4490 ± 166	20.0 ± 2.7	79 ± 5	2.27 ± 0.22	17.9 ± 0.6	4.46 ± 0.15	1.45 ± 0.10 18.1 ± 0.7		4.99 ± 0.46	1.60 ± 0.11 164 ± 8	
$2 - 4$	2004 - 1996	3854 ± 135	24.5 ± 2.5	96 ± 6	2.51 ± 0.25	19.2 ± 0.7	5.17 ± 0.21	1.52 ± 0.27 20.0 ± 0.8			4.79 ± 0.56 2.10 ± 0.15 146 ± 7	
$4 - 6$	1996 - 1988	3422 ± 126	15.8 ± 2.6	89 ± 5	2.65 ± 0.26	19.1 ± 0.7	4.23 ± 0.14	1.71 ± 0.11 19.7 ± 0.8			5.14 ± 0.45 1.88 ± 0.15 134 ± 7	
$6 - 8$	1988 - 1980	3126 ± 132	18.4 ± 2.6	89 ± 5	2.64 ± 0.26	19.8 ± 0.7	4.96 ± 0.20			1.76 ± 0.12 19.5 ± 0.8 5.08 ± 0.46 1.60 ± 0.11 123 ± 6		
$8 - 10$	1980 - 1972	3113 ± 204	18.1 ± 1.6	94 ± 6	2.93 ± 0.29	19.8 ± 0.7	4.57 ± 0.18	1.70 ± 0.11 19.7 ± 0.8			5.34 ± 0.48 1.74 ± 0.14 119 ± 6	
$10 - 12$	1972 - 1964	1490 ± 166	20.7 ± 1.5	88 ± 5	2.39 ± 0.24	20.4 ± 0.7	4.66 ± 0.18			1.83 ± 0.18 20.5 ± 0.8 3.47 ± 0.34 1.86 ± 0.15 120 ± 6		
$14 - 16$	1964 - 1961	2624 ± 111	24.9 ± 1.7	95 ± 5	1.69 ± 0.17	20.1 ± 0.7	4.49 ± 0.18	1.69 ± 0.11 19.9 ± 0.8			5.10 ± 0.59 2.02 ± 0.14 113 ± 6	
$16 - 18$	1961 - 1958	1033 ± 35	15.1 ± 1.7	93 ± 6	1.91 ± 0.17	20.0 ± 0.7	3.55 ± 0.11	1.77 ± 0.12 21.1 ± 0.8			5.39 ± 0.51 2.14 ± 0.18 123 ± 6	
$18 - 20$	1958 - 1954	1080 ± 34	17.5 ± 2.3	109 ± 9	1.44 ± 0.14	20.5 ± 0.7	4.60 ± 0.14	1.95 ± 0.15 20.9 ± 0.8			6.10 ± 0.53 1.80 ± 0.14 149 ± 9	
$20 - 22$	1954 - 1951	964 ± 34	24.7 ± 4.0	91 ± 6	1.49 ± 0.14	19.7 ± 0.7	3.86 ± 0.12			1.85 ± 0.11 21.6 ± 0.8 5.82 ± 0.42 2.37 ± 0.17 124 ± 6		
$22 - 24$	1951 - 1947	978 ± 38	16.0 ± 3.0	87 ± 6	1.52 ± 0.14	19.8 ± 0.7	3.59 ± 0.11	1.86 ± 0.12 21.6 ± 0.9			5.07 ± 0.49 2.35 ± 0.17 131 ± 6	
$24 - 26$	1947 - 1944	890 ± 29	20.0 ± 1.6	104 ± 8	1.23 ± 0.12	20.2 ± 0.7	3.88 ± 0.12	2.02 ± 0.16 21.3 ± 0.8			6.32 ± 0.71 2.42 ± 0.18 156 ± 9	
$26 - 28$	1944 - 1940	896 ± 30	13.9 ± 1.2	134 ± 10	0.84 ± 0.16	22.1 ± 0.8	3.30 ± 0.10	1.84 ± 0.14 22.4 ± 0.9			6.87 ± 0.77 2.48 \pm 0.20 104 \pm 6	
$28 - 30$	1940 - 1937	824 ± 28	22.9 ± 4.7	130 ± 8	0.53 ± 0.05	21.7 ± 0.8	3.77 ± 0.11	2.41 ± 0.30 22.8 ± 0.9			8.42 ± 0.60 2.56 ± 0.19 87 ± 4	
$30 - 32$	1937 - 1933	581 ± 32	16.4 ± 1.6	135 ± 8	0.74 ± 0.05	21.5 ± 0.8	3.32 ± 0.10	1.76 ± 0.12 22.9 ± 0.9			7.63 ± 0.60 2.94 ± 0.23	70 ± 4
$32 - 34$	1933 - 1930	705 ± 25	20.1 ± 1.9	133 ± 10	0.70 ± 0.07	21.4 ± 0.7	3.43 ± 0.11	1.94 ± 0.14 24.2 ± 0.9		7.30 ± 0.78 3.05 ± 0.23 90 ± 6		
$34 - 36$	1930 - 1926	938 ± 32	21.4 ± 3.0	117 ± 9	1.07 ± 0.10	20.9 ± 0.7	3.68 ± 0.12	2.12 ± 0.16 24.0 \pm 0.9			4.98 ± 0.41 2.97 ± 0.24 95 ± 7	
$36 - 40$	1926 - 1922	1048 ± 34	23.2 ± 4.0	96 ± 7	1.08 ± 0.10 21.1 ± 0.7		3.58 ± 0.11 2.07 ± 0.14 22.8 ± 0.9 4.81 ± 0.34 2.82 ± 0.20 77 ± 4					
	Continental Crust	25670	25.9	110	0.31	$\overline{7}$	4.7	1.5	10.3	2.5	1.5	52

275 **Table 5 – Enrichment Factors (EF) results (Only elements which showed along profile EF >1.5 or EF <0.8)**

Depth	Ca	Ce	Co	Fe	Hf	La	Lu	Na	Sb	Sm	$\mathbf U$	Yb	\mathbf{Zn}
$0 - 2$	0.8	1.4	1.5	1.6	0.8	1.5	0.8	5.0	2.6	1.5	1.2	0.7	2.5
$2 - 4$	0.8	1.3	1.3	1.4	0.8	1.4	0.9	4.0	2.5	1.6	1.1	0.8	2.1
$4 - 6$	0.9	$1.2\,$	1.2	1.4	0.8	1.3	0.8	3.6	2.7	1.3	1.2	0.7	1.9
$6 - 8$	0.8	1.2	1.2	1.3	0.7	1.3	0.7	3.2	2.6	1.5	1.1	0.6	1.7
$8 - 10$	0.6	1.2	1.1	1.3	0.7	1.4	0.7	3.2	2.9	1.4	1.2	0.7	1.6
$10 - 12$	0.6	1.1	1.0	1.2	0.7	1.1	0.7	1.5	2.3	1.3	0.7	0.7	1.6
$14 - 16$	0.8	1.2	0.9	1.1	0.8	1.2	0.8	2.6	1.6	1.3	1.1	0.8	1.5
$16 - 18$	0.9	1.3	1.0	1.1	0.9	0.9	0.7	1.0	1.9	1.0	1.2	0.8	1.7
$18 - 20$	0.8	1.4	1.0	1.1	0.8	1.0	0.8	1.1	1.4	1.3	1.3	0.7	2.0
$20 - 22$	0.8	1.5	$1.0\,$	1.2	0.9	1.1	0.8	1.0	1.5	1.2	1.3	0.9	1.7
$22 - 24$	0.7	1.4	1.2	1.3	0.9	1.1	0.8	$1.0\,$	1.5	1.1	1.1	0.9	1.8
$24 - 26$	0.7	1.4	1.1	1.2	0.9	0.9	0.9	0.9	1.2	1.1	1.4	0.9	2.1
$26 - 28$	0.7	1.6	0.9	1.0	0.8	0.7	0.8	0.8	0.7	0.9	1.4	0.8	1.3
$28 - 30$	0.6	1.8	1.0	1.0	0.8	0.9	0.9	0.8	0.5	1.0	1.7	0.9	1.1
$30 - 32$	0.6	1.9	1.0	1.0	0.8	1.0	0.8	0.5	0.7	0.9	1.6	$1.0\,$	0.9
$32 - 34$	0.7	1.7	1.2	1.1	1.0	0.9	0.9	0.7	0.6	0.9	1.5	1.1	1.2
$34 - 36$	0.7	1.5	1.2	1.1	1.1	0.8	1.0	0.9	1.0	1.0	1.0	1.1	1.2
$36 - 40$	1.0	1.0	1.0	1.0	1.0	1.0	1.0	1.0	1.0	1.0	1.0	1.0	1.0

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 Although *Igeo* results for Ce and U showed no indication of accumulation, the EF results 281 pointed to a discreet enrichment in the middle of the profile (Ce – 25-35 cm depth; $U - 25-30$ cm depth) (Figure 4). This could be related to sources of pollution which existed in the past but today either no longer exist or are under control. Figure 4 shows the behavior of these two elements along the profile.

Figure 4 – Behavior of Ce and U along the sediment profile

 In regards to the elements Na, Sb and Zn, EF results (> 2.0) at the top of the profile pointed to an enrichment which can be considered as anthropic origin, according to Sutherland (2000) or Zhang & Liu (2002) classification. The *Igeo* for these elements also presented values greater than zero. The concentrations for Na at the top of the profile were considered moderately enriched by both criteria (3.0 < EF < 5.0 and *Igeo* > 1.0), while 2.0 < EF < 3.0 and 0 < *Igeo <*1.0 for Sb and Zn were considered a light to moderate enrichment (Table 6).

 It should be noted that the values obtained in this study for various elements, most notably As, Cr, Ce, Fe, Zn, Sb, Th, Sc and U presented concentrations at the base of the profile (and in some cases throughout the entire profile) well above those set by Wedepohl (1995) as continental crust values. Concentrations obtained for Ba, Ca and Na at the reservoir were smaller than those established by Wedepohl (1995) (Table 4).

 It becomes clear that in this type of situation the use of local geochemical values for anthropic enrichment or contamination assessment is crucial for correct environmental evaluations. If in this study, the values set by Wedepohl (1995) had been taken as reference values, the conclusions would have been much different. For example, As which presented a concentration between 23.4 and 31.2 303 mg kg⁻¹ throughout the entire profile would have been considered highly contaminated with $Igeo > 3.0$ 304 and an EF >12.0 as compared to the defined value for As in the continental crust of 2.0 mg kg⁻¹. In this study the EF < 1.5 and *Igeo* < 0 were obtained for As using the concentration of the base of the profile as reference value.

 Other studies developed at this reservoir with the purpose of metal concentration evaluations were already published by Fávaro et al (2007), Bostelmann (2006) and Franklin et al (2012a). The concentration values obtained by these authors for REE and As, Ba, Co, Fe, Rb, Sb and Zn were very similar to the present study. These studies also concluded that the sediments from this reservoir, collected in points near the point of this study, could be contaminated for some elements such as As, Cs, Sb, Lu, Yb, Th, Hf and U due to the values obtained for the Enrichment Factors (EF > 2.0) by using the earth crust as reference values (Wedepohl, 1995). From this data it was possible to conclude that the choice of the background geochemical values is fundamental for the correct interpretation of the geochemical data as well as for the anthropic contamination evaluation. In the case of anthropic contamination evaluation of sediments, the utilization of a sedimentary profile can offer an adequate answer for this question.

3.4 – Metals determination by ICPOES, GF AAS and Direct Mercury Analyzer(DMA)

 The certified reference materials analyses for CRM050, SRM 8704, and SRM 2710 were done under the same conditions of the samples and the results obtained are shown in Table 7. The Z-score 322 was calculated for all results and were in the range of interval $|Z| < 3$ indicating good precision and accuracy for all analytical methodologies applied.

324 Tables 8 to 10 show the results obtained (mean of replicates \pm uncertainty) and the enrichment factor and *Igeo* for the elements analyzed, respectively. The uncertainty was calculated according to Franklin et al. (2012 b).

 Table 7 – Results for the Certified Reference Materials analyses by ICP OES, GF AAS and DMA (mg kg-1)

	Mn	Ni	Cu	C _d	Pb	Hg				
Certified Values										
CRM 050	636 ± 21	$344 + 6$	88.5 ± 1.7	80.0 ± 1.3	$111 + 2$	13.5 ± 0.4				
SRM 8704	544 ± 21	42.9 ± 3.7		2.94 ± 0.29	150 ± 17					
SRM 2710	2140 ± 60	8 ± 1	3420 ± 50	12.3 ± 0.3	5520 ± 30	9.88 ± 0.21				
Obtained Values										
CRM 050	593 ± 25	340 ± 9	89.1 ± 2.0	76.7 ± 2.2	110 ± 5	12.9 ± 1.1				
SRM 8704	479 ± 12	37.6 ± 1.4		3.45 ± 0.19	140 ± 6					
SRM 2710	1963 ± 80	7.21 ± 0.29	3355 ± 22	13.88 ± 0.57		9.52 ± 0.44				
Z score										
CRM 050	-1.32	-0.37	0.24	-1.29	-0.18	-0.51				
SRM 8704	-2.69	-1.34		1.47	-0.55					
SRM 2710	-1.77	-0.76	-1.19	2.45		-0.74				

330 **Table 8 – Results obtained for elements determined by ICPOES, GF AAS and DMA (mg kg-1** 331 **)**

Depth (cm)	Pb ^a	Hg^b	Cd ^a	Cu ^c	Ni ^c	Mn^c
$0 - 5$	52.9 ± 2.5	0.55 ± 0.08	1.02 ± 0.06	4867 ± 54	25.2 ± 2.1	907 ± 28
$5 - 10$	57.0 ± 2.4	0.68 ± 0.05	0.90 ± 0.07	5404 ± 62	26.4 ± 2.0	810 ± 21
$10 - 15$	69.7 ± 2.7	1.30 ± 0.10	0.69 ± 0.05	3774 ± 48	30.4 ± 2.1	599 ± 19
$15 - 20$	72.6 ± 2.9	2.69 ± 0.21	0.45 ± 0.03	1112 ± 27	33.9 ± 2.0	585 ± 22
$20 - 25$	60.0 ± 2.5	17.3 ± 1.1	0.30 ± 0.03	244 ± 10	22.9 ± 1.8	475 ± 16
$25 - 30$	42.4 ± 2.4	10.7 ± 0.9	0.06 ± 0.01	36.4 ± 1.9	14.9 ± 1.8	290 ± 13
$30 - 35$	38.6 ± 2.3	0.58 ± 0.06	0.05 ± 0.01	25.9 ± 1.1	13.3 ± 1.5	375 ± 15
$35 - 40$	29.0 ± 2.1	0.14 ± 0.03	0.04 ± 0.01	22.1 ± 1.0	10.9 ± 1.5	181 ± 12

332 **a – GF AAS; b – DMA; c – ICP OES**

334 **Table 9 – EF values obtained from metal results**

335

337 **Table 10 –** *Igeo* **values obtained from the metals results**

Geoaccumulation Index										
Depth (cm)	Pb	Hg	C _d	Cu	Ni	Mn				
$0 - 5$	0.3	1.4	4.1	7.2	0.6°	1.7				
$5 - 10$	0.4	1.7	3.9	7.3	0.7	1.6				
$10 - 15$	0.7	2.6	3.5	6.8	0.9	1.1				
$15 - 20$	0.7	3.7	2.9	5.1	1.1	1.1				
$20 - 25$	0.5	6.4	2.3	2.9	0.5	0.8				
$25 - 30$	0.0	5.7	0.0	0.1	< 0	0.1				
$30 - 35$	< 0	1.5	< 0	< 0	< 0	0.5				
$35 - 40$	< 0	< 0	< 0	< 0	< 0	< 0				

 It is clear that the great enrichment (EF) for Cd and Cu in the top of the profile, with values of 26.9 and 232 respectively, confirmed by *Igeo* of 4.1 and 7.2 respectively, indicates extremely contaminated sediment (Tables 9 and 10). Such surface contamination for these elements indicates that the probable sources may still be active in the reservoir and may still be releasing such elements. In the case of Cu, this may be due to the use of their salts to prevent algae growth in the reservoir. Regarding Ni and Mn a moderate enrichment more significant on the surface of the profile could be observed. This moderate contamination was also confirmed by *Igeo*, with values between 0 and 1 for Ni and 1.1 and 1.7 for Mn (Table 10). In relation to Hg and Pb enrichment this is more significant in the middle of the profile, indicating that in the past there was a strong anthropic contribution of these elements, especially for Hg (EF from 91 to 128 and *Igeo* from 5.7 to 6.4). The current depletion of these concentrations in the surface is an indication that these sources are controlled or no longer exist. Figure 5 shows the behavior of Cd, Hg and Pb along the sediment profile.

 Fávaro et al. (2007) determined the multielemental concentrations of sediment cores from Rio 352 Grande reservoir by INAA, Hg content by CV AAS and sedimentation rates by ²¹⁰ Pb method. For the sediment core located near the catchment point of the water supply (near the sediment profile collected 354 in the present study), a peak of 11,586 mg Hg kg^{-1} was found in 1949 and at the end of the core 355 (1932), values about 0.145 mg Hg kg^{-1} were found. In the present study a peak in the Hg content of 356 17.3 mg kg⁻¹ was found around the 1950's (Figure 5) decreasing to 0.14 mg kg⁻¹ at the 35-40 cm depth (approximately 1920). The results obtained for Hg were very similar in both studies.

Figure 5 – Behavior of Pb, Hg and Cd along the sediment profile

4 – CONCLUSION

363 The ²¹⁰Pb profile showed an average accumulation rate of 0.26 cm y⁻¹ in the upper segments of 364 the sediment profile (2011 to 1964 year) and average accumulation rate of 0.57 cm y^{-1} (1964 to 1922 year) in the lower section, clearly reflecting the different periods of sediment loading in this reservoir.

 The Enrichment Factor and Geoaccumulation Index are excellent tools for environmental evaluations and in this study presented results aligned with the anthropic enrichment of some of the elements analyzed.

 Most of the elements presented concentrations very similar over time in this reservoir. Cu and Cd have severe anthropic contributions in the reservoir, the same holding true for Hg. But for this element a peak in concentration in the middle of the profile was observed, indicating that there was a source in the past but nowadays appears to be under control, although the concentration in the surface 373 is still considered high $(0.55 \text{ mg kg}^{-1})$. The reservoir still presents a discreet anthropic enrichment for Ni, Pb and Sb and a moderate enrichment for Na and Mn. All other elements and rare earth elements, in general, presented constant values throughout the sediment profile.

 For this study background values for metals and trace elements in the sediments from the Rio Grande Reservoir were considered as the concentration present in the lowest slice of the profile corresponding to the age of 1920. At that time the industrialization of São Paulo was nearly non- existent and in its very early development stages. From the data of this study it can be seen that the choice of the local geochemical background values are fundamental for correct interpretations of anthropic contamination evaluation of sediments.

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